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# Comparison of metoprolol degradation by Fe<sup>III</sup>-NTA modified Fenton-like reaction in the absence and presence of manganese: Efficiency and intermediates



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#### HIGHLIGHTS

- Metoprolol degradation by Fe<sup>III</sup>-NTA/ H<sub>2</sub>O<sub>2</sub> was enhanced in the presence of Mn<sup>II</sup>.
- Rapid degradation of metoprolol was obtained by Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn at pH 4.0-8.0.
- Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn led to more efficient degradation of metoprolol and intermediates.
- Degradation pathways of metoprolol were proposed.

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#### GRAPHICAL ABSTRACT



#### ABSTRACT

The degradation of metoprolol in aqueous solutions by traditional Fenton-like reaction (Fe<sup>III</sup>/H<sub>2</sub>O<sub>2</sub>), and Fe<sup>III</sup>-NTA modified Fenton-like reaction (Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>) in the absence and presence of Mn<sup>II</sup> (Fe<sup>III</sup>-NTA/ H<sub>2</sub>O<sub>2</sub>-Mn) have been investigated. The results show that Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> is able to degrade metoprolol at initial neutral pH. In particular, the presence of Mn<sup>II</sup> greatly improved the degradation rate of metoprolol by Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> over a wide pH range of 4.0–8.0. Under the same conditions, the degradation rate constants of metoprolol (k) obtained in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system were typically 7–9-fold larger than those obtained in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system. The involved reactions in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system were proposed on the basis of important parameters analysis including ferrous ion concentration, dissolved oxygen (DO) concentration, and the quenching experiments for hydroxyl radial (HO·) and superoxide anion radical (O<sub>2</sub><sup>-</sup>). Several intermediates have been identified by mass spectrometry. Our results suggest that the degradation of metoprolol in both Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> and Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn systems were caused by HO attack. The degradation pathways of metoprolol were proposed on the basis of the identified intermediates. Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system led to more efficient degradation of metoprolol and its intermediates. © 2016 Elsevier B.V. All rights reserved.

#### 1. Introduction

Metoprolol, a selective β1 receptor blocking drug, is widely used in the therapy of hypertension and angina. In the past decade, metoprolol and its metabolites have been detected in sewage treatment plant effluents, surface water, groundwater, and even

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in drinking water, which pose a potential risk to human health [1,2]. The degradation rate of metoprolol in water by the direct photolysis is small, e.g., the extrapolated half-lives of metoprolol in central Europe and the U.S. were reported at 28–95 d in summer and 190–449 d in winter [3]. Thus, metoprolol is most likely to be pseudo-persistent in natural waters if do not consider the indirect photolysis effects. In addition, the efficiency of metoprolol removal is lower than 30% in traditional sewage treatment plant [4]. There is an urgent need to improve the removal efficiencies of pharmaceuticals in water.

A far more effective method is the so-called advanced oxidation processes (AOPs), involving the generation of hydroxyl radical (HO<sup>•</sup>) [5–7], have been widely applied in water treatment for harmful organic compounds removal. Efficient degradation of metoprolol has been obtained by photocatalysis, UV/H<sub>2</sub>O<sub>2</sub> and Fenton-related processes [8–10]. Two decades ago, research by Sun and Pignatello [11] showed the successful degradation of 2,4-dichlorophen-oxyacetic acid at pH 6.0 by chelates modified Fenton-like reaction. Chelating agents modified Fenton-like reaction has been receiving increasing attention because it can work at neutral pH conditions, which overcomes the limitation of acidic pH (e.g., optimal pH 2.8-3.0) by using traditional Fenton and/or Fenton-like reaction. S,S-ethylenediamine-N,N'-disuccinic acid (EDDS) modified Fenton reaction (i.e., Fe<sup>II</sup>-EDDS/H<sub>2</sub>O<sub>2</sub>) and Fenton-like reaction (i.e., Fe<sup>III</sup>-EDDS/H<sub>2</sub>O<sub>2</sub>) are capable of degrading 4-chlorophenol, bisphenol A, naphthenic acid, and 4-tertbutylphenol at near neutral or basic pH conditions [12-15]. The use of Fe<sup>III</sup>-EDDS for photo-Fenton reaction at neutral pH is also successful in degrading micro-pollutants in wastewater treatment plant effluents [16–18]. In addition, nitrilotriacetic acid (NTA) and ethylenediaminetetraacetic acid (EDTA) modified Fenton-like reactions (i.e., Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> and Fe<sup>III</sup>-EDTA/H<sub>2</sub>O<sub>2</sub>) have also been proved to be highly effective in degrading atrazine, carbamazepine, fenuron, naphthenic acid, parachlorobenzoic acid and sulfamethoxazole at neutral or slightly basic pH conditions [19–22]. In particular, our recent finding suggested that the presence of manganese (Mn<sup>II</sup>) could greatly enhance the Fenton-like catalytic activity of Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> [23]. The potential mechanism may involve an enhanced generation of superoxide anion radical  $(O_2^{\cdot-})$  by the Mn<sup>II</sup>, which can accelerate the reduction of Fe<sup>III</sup>-NTA to Fe<sup>II</sup>-NTA and indirectly improve the generation of HO [23]. The Mn<sup>II</sup>/Fe<sup>III</sup>-NTA catalytic system would be very promising since Fe<sup>III</sup> and Mn<sup>II</sup> are quite ubiquitous in the environment and can be used as natural Fenton-like catalyst in the presence of NTA (or natural Fe<sup>III</sup> complexing agents) and H<sub>2</sub>O<sub>2</sub>. However, little research on this system has been reported.

The purpose of this study is to demonstrate the potential effectiveness of Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn for water treatment. In the present study, we choose metoprolol as a target compound, to evaluate the efficiency and intermediates generated from metoprolol degradation by Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> in the absence and presence of Mn<sup>II</sup>. The degradation kinetics of metoprolol in both systems were compared under various conditions. The involved reactions in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system were proposed on the basis of certain important parameters analysis such as ferrous ion concentration, dissolved oxygen (DO) concentration, and the quenching experiments for HO<sup>-</sup> and O<sub>2</sub><sup>--</sup>. The intermediates were identified by liquid chromatography/mass spectrometry.

#### 2. Materials and methods

#### 2.1. Chemicals

Ferric chloride hexahydrate, hydrogen peroxide  $(30\% H_2O_2, w/w)$ , sodium hydroxide and sulfuric acid were purchased from Sino-

pharm Chemical Reagent Co., Ltd. Metoprolol tartrate (purity  $\ge$  98%) was purchased from LKT Laboratories, Inc. NTA was purchased from Acros Organics. Formic acid and manganese nitrate tetrahydrate were purchased from Sigma-Aldrich. HPLC-grade methanol, 2-propanol and water were purchased from Fisher Scientific. Chloroform was purchased from Chinasun Specialty Products Co., Ltd. All chemicals were of reagent grade or better, and used as received without further purification. Milli-Q water was used to prepare solutions.

#### 2.2. Experimental procedures

All experiments were performed in 125 ml glass Erlenmeyer flasks. Briefly, 100 ml of metoprolol aqueous solution and an appropriate amount of freshly prepared Fe<sup>III</sup>-NTA solution were added into the flask. The flask was then placed in a waterjacketed glass vessel water bath  $(25 \pm 1 \text{ °C})$  under a magnetic stirring condition. After that, an appropriate amount of Mn<sup>II</sup> was added to the solution. The pH of the solution was then adjusted to the desired value using 1.0/0.1 M NaOH or 1.0/0.1 M H<sub>2</sub>SO<sub>4</sub>. Finally, Fenton-like reaction was started by the addition of an appropriate amount of H<sub>2</sub>O<sub>2</sub>. It is worth noting that the waterjacketed glass vessel was covered with aluminum foil to prevent the effect of light on the Fenton-like reaction. Samples (1 ml) were withdrawn at different time intervals and mixed with 0.1 ml of methanol immediately to quench the reaction.

#### 2.3. Analytical methods

The concentration of metoprolol was determined by a high performance liquid chromatography (HPLC, Agilent 1100) with fluorescence detection. An Ultra C-18 column (5  $\mu$ m, 250  $\times$  4.6 mm) was used for the chromatographic separation. The mobile phase consisted of methanol and 0.1% formic acid aqueous solution (35:65, v/v) at a flow rate of 0.9 ml min<sup>-1</sup>. The column temperature was 30 °C. The injection volume was 20  $\mu$ l, and the fluorescence detection was performed at 275 nm of excitation and 298 nm of emission.

The degradation intermediates were identified by an Agilent 1200 HPLC system coupled to a 6130 quadrupole mass spectrometer. The MS detector was performed for a full scan range of m/z 30–600 under an ESI + APCI positive mode. The MS operating conditions were as follows: Fragmentor 90, drying gas flow rate 6 L min<sup>-1</sup>, nebulizer pressure 45 psi, dry gas temperature 300 °C, vaporizer temperature 250 °C, capillary voltage 3500 V, corona current 4.0 µA and charging voltage 2000 V.

The concentration of ferrous ion in the aqueous solutions was determined by using the 1,10-phenathroline method. The concentration of  $H_2O_2$  was determined by the iodide method [24]. The dissolved oxygen (DO) concentration in the aqueous solutions during the reaction time was monitored by a  $dO_2$  sensor (Model Z010023525, AppliSens).

#### 3. Results and discussion

3.1. Comparison of metoprolol degradation by  $Fe^{III}/H_2O_2$ ,  $Fe^{III}$ -NTA/ $H_2O_2$  and  $Fe^{III}$ -NTA/ $H_2O_2$ -Mn

Fig. 1 shows a comparison of metoprolol degradation by  $Fe^{III}/H_2O_2$ ,  $Fe^{III}-NTA/H_2O_2$  and  $Fe^{III}-NTA/H_2O_2$ -Mn systems at initial neutral pH. A minimum degradation of metoprolol was observed in  $Fe^{III}/H_2O_2$  system, with 3.61% degradation efficiency of metoprolol after 60 min of the reaction time. This result was expected because of the precipitation of ferric ion at pH 7.0 in the absence of chelator. In contrast,  $Fe^{III}$ -NTA complex is able to catalyze



**Fig. 1.** Comparison of metoprolol degradation by Fe<sup>III</sup>/H<sub>2</sub>O<sub>2</sub>, Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> and Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn systems. Experimental conditions: 29.2  $\mu$ M of metoprolol, 0.1 mM of Fe<sup>III</sup>, NTA/Fe<sup>III</sup> molar ratio of 2:1, H<sub>2</sub>O<sub>2</sub>/Fe<sup>III</sup> molar ratio of 100:1, Mn<sup>II</sup>/Fe<sup>III</sup> molar ratio of 1:1, and initial pH of 7.0.

Fenton-like reaction at pH 7.0, resulted in 100% degradation efficiency of metoprolol after 30 min of the reaction time in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system. In particular, Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system led to a very fast degradation of metoprolol, with 100% degradation efficiency of metoprolol after 7 min of the reaction time. The result is consistent with our recent finding that the presence of Mn<sup>II</sup> could enhance Fe<sup>III</sup>-NTA-catalyzed Fenton-like reaction [23].

The effects of Mn<sup>II</sup> concentration on the degradation kinetics of metoprolol in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system were shown in Fig. 2. An increase in Mn<sup>II</sup> concentration from 0.01 to 0.1 mM, corresponding to Mn:Fe molar ratios of 0.1:1 to 1:1, had a positive effect on metoprolol degradation. The apparent degradation rate constant of metoprolol (k) increased from 0.7547 to 1.6048 min<sup>-1</sup>. However, further increasing Mn<sup>II</sup> concentration from 0.2 to 1.0 mM led to a slight inhibitory effect on the degradation rate of metoprolol, e.g., the *k* value of metoprolol decreased from 1.6289 to 1.1103 min<sup>-1</sup>. This is because when Mn<sup>II</sup> concentration was larger than 0.2 mM, hydrous manganese dioxide could form from free Mn<sup>II</sup> in excess under the given conditions, which is capable of catalyzing the decomposition of  $H_2O_2$  to  $O_2$  and  $H_2O$  [25]. It is worth noting that even at a high concentration of 1.0 mM of Mn<sup>II</sup>, corresponding to Mn:Fe molar ratio of 10:1, the k value of metoprolol was still larger (6-fold) than that in the absence of Mn<sup>II</sup>.

Quenching tests showed that the degradation of metoprolol in  $Fe^{III}$ -NTA/H<sub>2</sub>O<sub>2</sub>-Mn system was significantly inhibited by the addi-



**Fig. 2.** Effects of  $Mn^{II}$  concentration on the degradation kinetics of metoprolol in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system. Experimental conditions: 29.2  $\mu$ M of metoprolol, 0.1 mM of Fe<sup>III</sup>, NTA/Fe<sup>III</sup> molar ratio of 2:1, H<sub>2</sub>O<sub>2</sub>/Fe<sup>III</sup> molar ratio of 100:1, Mn<sup>II</sup>/Fe<sup>III</sup> molar ratio of 0.1:1–10:1, and initial pH of 7.0.

tion of 2-propanol, but not by chloroform (Fig. S1, in Supplementary data), which suggests that the degradation of metoprolol was probably caused by HO<sup>•</sup> attack. The presence of Mn<sup>II</sup> is unable to directly catalyze Fenton-like reaction to generate HO; this is because no degradation of metoprolol was observed in the absence of Fe<sup>III</sup> (e.g., Mn<sup>II</sup>-NTA/H<sub>2</sub>O<sub>2</sub> in Fig. 1). Thus, the very fast degradation of metoprolol in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system could be attributed to an indirect enhancement of Fe<sup>III</sup>-NTA-catalyzed Fenton-like reaction by Mn<sup>II</sup>. One hypothesis is that the presence of Mn<sup>II</sup> favors the production of O<sub>2</sub><sup>.-</sup>, which enhances the reduction of Fe<sup>III</sup>-NTA to Fe<sup>II</sup>-NTA, and in turn increases the production of HO<sup>·</sup> from the reaction of Fe<sup>II</sup>-NTA with H<sub>2</sub>O<sub>2</sub>. In order to prove this, the concentration of ferrous ion in the aqueous solutions was monitored during the reaction time, and the results were shown in Fig. 3. It can be seen that the concentration of ferrous ion in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system increased rapidly from 0 to  $0.40 \text{ mg L}^{-1}$  within 5 min of reaction time, and then gradually decreased to 0.11 mg  $L^{-1}$  within 30 min of reaction time. On the contrary, the concentration of ferrous ion in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system was always below 0.02 mg  $L^{-1}$ , which was much smaller than that in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system. In addition, changes in DO concentrations in both systems with the reaction time were also monitored. It can be observed from Fig. 3 that the DO concentration in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system increased significantly from 100% (i.e., 100% air saturation) to 275% within 7 min of reaction time, and then gradually decreased to 211% within 30 min of reaction time, which was much higher than that in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system. The results suggested that more oxygen was produced in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system. Moreover, Fig. 3 provides further evidence that the decomposition rate of H<sub>2</sub>O<sub>2</sub> in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system was faster than that in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system.

On the basis of Haber-Weiss reactions, the involved main reactions in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system were proposed by R1-R8. As shown in reactions R1-R3, it was O<sub>2</sub><sup>--</sup> and not free HO which was generated from the reaction of Mn<sup>II</sup>-NTA with H<sub>2</sub>O<sub>2</sub>. One the basis of reactions R4 and R5, O<sub>2</sub><sup>--</sup> should rapidly reduce Fe<sup>III</sup>-NTA (and Mn<sup>III</sup>-NTA) to Fe<sup>II</sup>-NTA (and Mn<sup>II</sup>-NTA), leading to the production of oxygen. Thus, R1-R5 also explain the rapid decomposition of H<sub>2</sub>O<sub>2</sub> and the significant increase of DO concentration in Fe<sup>III</sup>-NTA/ H<sub>2</sub>O<sub>2</sub>-Mn system. Most importantly, HO is generated from the reaction of Fe<sup>II</sup>-NTA with H<sub>2</sub>O<sub>2</sub>, which lead to metoprolol destruction (as shown in R6 and R8). Because Fe<sup>II</sup>-NTA is mainly generated



Fig. 3. Changes in DO concentration,  $H_2O_2$  concentration and ferrous ion concentration with the reaction time in  $Fe^{III}$ -NTA/ $H_2O_2$  and  $Fe^{III}$ -NTA/ $H_2O_2$ -Mn systems. Experimental conditions: 29.2  $\mu$ M of metoprolol, 0.1 mM of  $Fe^{III}$ , NTA/ $Fe^{III}$  molar ratio of 2:1,  $H_2O_2/Fe^{III}$  molar ratio of 100:1,  $Mn^{II}/Fe^{III}$  molar ratio of 1:1, and initial pH of 7.0.

from the reaction of Fe<sup>III</sup>-NTA with  $H_2O_2$  in Fe<sup>III</sup>-NTA/ $H_2O_2$  system (i.e., R7), thus the generation rate of Fe<sup>II</sup>-NTA in Fe<sup>III</sup>-NTA/ $H_2O_2$  system is much slower than that in Fe<sup>III</sup>-NTA/ $H_2O_2$ -Mn system.

$$Mn^{II}-NTA + 2H_2O_2 \leftrightarrow Mn^{II}-NTA - 2H_2O_2$$
(R1)

 $Mn^{II}\text{-}NTA\text{-}2H_2O_2 \rightarrow Mn^{III}\text{-}NTA\text{-}H_2O_2\text{-}HO^{\text{-}} + OH^{-} \tag{R2}$ 

$$Mn^{\text{III}}\text{-NTA-H}_2O_2\text{-HO}^{\cdot} \rightarrow Mn^{\text{III}}\text{-NTA} + O_2^{\cdot-} + H_2O + H^+$$
(R3)

 $Mn^{III}\text{-}NTA + O_2^{-} \rightarrow Mn^{II}\text{-}NTA + O_2 \tag{R4}$ 

 $Fe^{III}-NTA + O_2^{-} \xrightarrow{Fast} Fe^{II}-NTA + O_2$ (R5)

 $Fe^{II}\text{-}NTA + H_2O_2 \rightarrow Fe^{III}\text{-}NTA + HO^{\text{-}} + OH^{-} \tag{R6}$ 

 $Fe^{III}-NTA + H_2O_2 \xrightarrow{Slow} Fe^{II}-NTA + 2H^+ + O_2^{--}$ (R7)

 $Metoprolol + HO' \rightarrow Degradation \ products \tag{R8}$ 

#### 3.2. Effects of initial pH

The effects of initial pH on the degradation kinetics of metoprolol by Fe<sup>III</sup>/H<sub>2</sub>O<sub>2</sub>, Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> and Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn systems were investigated over the range of pH 3.0-9.0 (Fig. 4). In Fe<sup>III</sup>/H<sub>2</sub>O<sub>2</sub> system, a k value of 0.0976 min<sup>-1</sup> was obtained at an initial pH 3.0. When the initial pH value was higher than 5.0, the degradation of metoprolol was significantly inhibited in Fe<sup>III</sup>/H<sub>2</sub>O<sub>2</sub> system due to the precipitation of ferric ion. In Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system, an increase in the initial pH value from 3.0 to 9.0 had a positive effect on the degradation rate of metoprolol, resulted in increasing the k value of metoprolol from 0.0033 to 0.7190 min<sup>-1</sup>. The small degradation rates of metoprolol under acidic pH conditions (e.g., pH 3.0-4.0) can be explained either by relatively low regeneration rate of Fe<sup>II</sup>-NTA, or by low catalytic activity of Fe<sup>III</sup>-NTA species toward H<sub>2</sub>O<sub>2</sub> at acidic pH. First of all, O<sub>2</sub><sup>.-</sup> was believed to play an important role in reducing Fe<sup>III</sup>-NTA to Fe<sup>II</sup>-NTA which then react with H<sub>2</sub>O<sub>2</sub> to generate HO<sup>•</sup>. Since more  $O_2^{\cdot-}$  changes to perhydroxyl radical (HOO<sup>•</sup>) at pH 3.0–4.0 (Eq. (1)) [26], the regeneration rate of Fe<sup>II</sup>-NTA can be limited due to the weak reduction capability of HOO'.

$$HOO^{\cdot} \leftrightarrow O_2^{\cdot -} + H^+ \quad pKa = 4.8 \tag{1}$$

On the other hand, Fe<sup>III</sup>-NTA species and free NTA also react with HO<sup>•</sup>. The reaction between HO<sup>•</sup> and Fe<sup>III</sup>-NTA species/free



**Fig. 4.** Effects of initial pH on the degradation kinetics of metoprolol by Fe<sup>III</sup>/H<sub>2</sub>O<sub>2</sub>, Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> and Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn systems. Experimental conditions: 29.2  $\mu$ M of metoprolol, 0.1 mM of Fe<sup>III</sup>, NTA/Fe<sup>III</sup> molar ratio of 2:1, H<sub>2</sub>O<sub>2</sub>/Fe<sup>III</sup> molar ratio of 100:1, Mn<sup>II</sup>/Fe<sup>III</sup> molar ratio of 1:1, and initial pH of 3.0–9.0.

NTA would have an inhibiting effect on the degradation rate of metoprolol since the reactions are competing. A larger reaction rate between HO and Fe<sup>III</sup>-NTA species/free NTA would lead to a smaller degradation rate of metoprolol. According to the literature [21,27], Fe<sup>III</sup>-NTA species and free NTA are strongly pH-dependent. Although the reaction rate constant of Fe<sup>III</sup>-NTA species with HO<sup>•</sup> is rarely reported (e.g.,  $1.6 \times 10^8 \text{ M}^{-1} \text{ s}^{-1}$  was reported at pH 2.0 [28]), the reaction rate constants of HO with free NTA have been reported at  $6.1 \times 10^7$ ,  $7.5 \times 10^8$ ,  $5.5 \times 10^8$ ,  $4.77 \times 10^8$ ,  $2.5 \times 10^9$ and  $4.2 \times 10^9 \,\mathrm{M^{-1} \, s^{-1}}$  at pH 2.0, 4.0, 6.0, 8.0, 9.0 and 10 [21,27], respectively. It can be seen that the reaction rate of HO<sup>•</sup> with free NTA was increased with the increase of pH from 2.0 to 10. Thus, the decrease of pH should favor the degradation of metoprolol. However, the present and previous studies have shown that the fast degradation of model contaminants and the decomposition of H<sub>2</sub>O<sub>2</sub> occurred at neutral or slight basic pH in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system [20.27], which rules out the effect of free NTA. Therefore, it is logically possible that Fe<sup>III</sup>-NTA species at acidic pH (i.e., Fe<sup>III</sup>(NTA)  $(H_2O)_2$  [29]) might be inefficient for the activation of  $H_2O_2$  to induce Fenton-like reaction, or the reaction of Fe<sup>III</sup>(NTA)(H<sub>2</sub>O)<sub>2</sub> with HO<sup>•</sup> might act as an inhibitory reaction.

Maximum degradation rates of metoprolol were obtained in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system. It can be seen that the *k* values of metoprolol were much larger than those in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system over the initial pH range of 4.0–8.0. At initial pH 3.0, the k value of metoprolol was 0.0819 min<sup>-1</sup> in Fe<sup>III</sup>-NTA/ $H_2O_2$ -Mn system, which was close to that in  $Fe^{III}/H_2O_2$  system (0.0976 min<sup>-1</sup>). With increasing the initial pH from 3.0 to 5.0, the k values of metoprolol increased greatly from 0.0819 to 2.5809 min<sup>-1</sup> in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system. On the basis of the pKa value of HOO<sup>·</sup>/O<sub>2</sub>. (pKa = 4.8) [26], the fraction of O<sub>2</sub><sup>-</sup> was calculated at 1.56%, 13.68% and 61.31% (of the total amount of HOO and  $O_2$ ) at pH 3.0, 4.0 and 5.0, respectively. It can be seen that the increase of the *k* values of metoprolol was consistent with the increase of the fraction of  $O_2$ .<sup>-</sup> at pH 3.0–5.0. Thus, the results support our hypothesis that O<sub>2</sub><sup>--</sup> plays an important role in reducing Fe<sup>III</sup>-NTA to Fe<sup>II</sup>-NTA, which in turn improves the generation of HO<sup>•</sup> and leads to a rapid degradation of metoprolol. It is worth noting that the fraction of HOO<sup>•</sup> is 38.69% at pH 5.0. HOO<sup>•</sup> is also an oxidant with an oxidation potential of 1.78 V [30], weaker than HO; thus the fastest degradation of metoprolol at pH 5.0 can be explained by HOO' along with HO' attack. At neutral pH, the fraction of  $O_2$ <sup>--</sup> would increase to >99% and the fraction of HOO<sup>-</sup> would decrease to <1%. Therefore, the degradation of metoprolol at neutral pH would be mainly caused by HO<sup>•</sup> attack. The large fraction of O<sub>2</sub><sup>.-</sup> at neural pH or slight basic pH could favor the generation of Fe<sup>II</sup>-NTA. Unlike Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system, however, the k values of metoprolol in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system gradually decreased with the increasing initial pH from 5.0 to 9.0. In particular, at initial pH 9.0, the k value of metoprolol in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system was smaller than that in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system. This can be explained by the fact that the complexation constants of NTA with Fe<sup>III</sup>/Fe<sup>II</sup>(log k = 15.9 and 9.8) are much larger than that with  $Mn^{II}$  (log k = 7.46) [31], thus Mn<sup>II</sup> would firstly leach due to the degradation of NTA by HO attack. The leached Mn<sup>II</sup> could form hydrous manganese dioxide when the solution pH value is >5.0, which can directly catalyze the decomposition of  $H_2O_2$  to  $O_2$  and  $H_2O$  [25]. Additionally, the decrease of Mn<sup>II</sup> concentration in aqueous phase would lead to the rate of  $O_2$ .<sup>-</sup> production decreasing, which in turn decrease the generation rates of Fe<sup>II</sup>-NTA and HO<sup>.</sup>

#### 3.3. Effects of NTA:Fe molar ratios

The effects of NTA/Fe molar ratios on the degradation kinetics of metoprolol by  $Fe^{III}$ -NTA/H<sub>2</sub>O<sub>2</sub> and  $Fe^{III}$ -NTA/H<sub>2</sub>O<sub>2</sub>-Mn systems at initial neutral pH were investigated (Fig. 5). It can be seen that over

the range of NTA:Fe molar ratios tested, the *k* values of metoprolol in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system were 7–17-fold larger than those in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system. An increase in NTA/Fe molar ratio from 0.5:1 to 2:1 had positive effects on the degradation rates of metoprolol in both Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> and Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn systems. The k values of metoprolol increased from 0.0311 to 0.1754 min<sup>-1</sup> in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system and increased from 0.3450 to 1.6048 min<sup>-1</sup> in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system, respectively. When NTA/Fe molar ratio was larger than 2:1, however, a slight decrease in the *k* values of metoprolol was observed in either Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> or Fe<sup>III</sup>-NTA/  $H_2O_2$ -Mn system. This can be explained by the fact that the free NTA in excess scavenges HO. Although the reaction rate constant of metoprolol with HO (5.2–8.4  $\times$  10<sup>9</sup> M<sup>-1</sup> s<sup>-1</sup>) [32,33] is larger than that of Fe<sup>III</sup>-NTA/free NTA with HO<sup>•</sup>, it is believed that the competitive consumption of HO<sup>•</sup> by the free NTA in excess can slightly limit the degradation of metoprolol.

#### 3.4. Effects of $H_2O_2$ : Fe<sup>III</sup> molar ratios

Fig. 6 shows the effects of  $H_2O_2/Fe^{III}$  molar ratios on the degradation kinetics of metoprolol by Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> and Fe<sup>III</sup>-NTA/ H<sub>2</sub>O<sub>2</sub>-Mn systems at initial neutral pH. Without the addition of H<sub>2</sub>O<sub>2</sub>, no degradation of metoprolol was observed in both systems. Over the range of  $H_2O_2/Fe^{III}$  molar ratios tested, the k values of metoprolol in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system were 5–9-fold larger than those in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system. With increasing H<sub>2</sub>O<sub>2</sub>/Fe<sup>III</sup> molar ratio from 10:1 to 200:1, the *k* values of metoprolol continuously increased from 0.0598 to 0.2276 min<sup>-1</sup> in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system. In Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system, the k values of metoprolol increased greatly from 0.3384 to 1.6048 min<sup>-1</sup> and reached a plateau with increasing H<sub>2</sub>O<sub>2</sub>/Fe<sup>III</sup> molar ratio from 10:1 to 100:1. It is expected that the use of higher molar ratio of H<sub>2</sub>O<sub>2</sub>/Fe<sup>III</sup> led to higher rate of HO generation, by which faster degradation of metoprolol could be obtained. No inhibitory effect of H<sub>2</sub>O<sub>2</sub> on the degradation kinetics of metoprolol was observed even at H2O2/Fe<sup>III</sup> molar ratio of 200:1. This is because the reaction rate constant of  $H_2O_2$  with HO' ((1.2–4.5)  $\times$   $10^7\,M^{-1}\,s^{-1})$  [34] is much smaller than those of metoprolol, Fe<sup>III</sup>-NTA and free NTA with HO.

### 3.5. Effect of Fe<sup>III</sup> dosage

 $Fe^{III}$  acts as the catalyst to initiate Fenton-like reaction, a high rate of HO<sup>o</sup> generation might be obtained at high concentration of Fe<sup>III</sup>-NTA. The effects of Fe<sup>III</sup> dosages on the degradation kinetics of metoprolol by Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> and Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn systems at



**Fig. 5.** Effects of NTA/Fe<sup>III</sup> molar ratios on the degradation kinetics of metoprolol by Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> and Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn systems. Experimental conditions: 29.2  $\mu$ M of metoprolol, 0.1 mM of Fe<sup>III</sup>, NTA/Fe<sup>III</sup> molar ratio of 0.5:1-5:1, H<sub>2</sub>O<sub>2</sub>/Fe<sup>III</sup> molar ratio of 100:1, Mn<sup>II</sup>/Fe<sup>III</sup> molar ratio of 1:1, and initial pH of 7.0.



**Fig. 6.** Effects of  $H_2O_2/Fe^{III}$  molar ratios on the degradation kinetics of metoprolol by  $Fe^{III}$ -NTA/ $H_2O_2$  and  $Fe^{III}$ -NTA/ $H_2O_2$ -Mn systems. Experimental conditions: 29.2  $\mu$ M of metoprolol, 0.1 mM of  $Fe^{III}$ , NTA/ $Fe^{III}$  molar ratio of 2:1,  $H_2O_2/Fe^{III}$  molar ratio of 5:1–200:1,  $Mn^{II}/Fe^{III}$  molar ratio of 1:1, and initial pH of 7.0.



**Fig. 7.** Effects of Fe<sup>III</sup> dosages on the degradation kinetics of metoprolol by Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> and Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn systems. Experimental conditions: 29.2  $\mu$ M of metoprolol, 0.01–0.15 mM of Fe<sup>III</sup>, NTA/Fe<sup>III</sup> molar ratio of 2:1, H<sub>2</sub>O<sub>2</sub>/Fe<sup>III</sup> molar ratio of 100:1, Mn<sup>II</sup>/Fe<sup>III</sup> molar ratio of 1:1, and initial pH of 7.0.



**Fig. 8.** Changes in the total molecular ion chromatogram area of metoprolol and its intermediates with the reaction time. Experimental conditions: 292  $\mu$ M of metoprolol, 0.1 mM of Fe<sup>III</sup>, NTA/Fe<sup>III</sup> molar ratio of 2:1, H<sub>2</sub>O<sub>2</sub>/Fe<sup>III</sup> molar ratio of 100:1, Mn<sup>II</sup>/Fe<sup>III</sup> molar ratio of 1:1, and initial pH of 7.0.

initial neutral pH were shown in Fig. 7. No degradation of metoprolol was observed in the absence of Fe<sup>III</sup>, which suggests that HO<sup> $\cdot$ </sup> is generated from the Fenton-like reaction of Fe<sup>III/II</sup>-NTA/H<sub>2</sub>O<sub>2</sub>. Over



## Further degradation products

Scheme 1. Intermediates and degradation pathways of metoprolol.

the range of Fe<sup>III</sup> concentrations tested, the *k* values of metoprolol in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system were 7–9-fold larger than that in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system. An increase in Fe<sup>III</sup> concentrations from 0 to 0.15 mM has positive effects on the degradation rate of metoprolol in both systems. The *k* values of metoprolol increased from 0 to 0.3174 min<sup>-1</sup> in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system and to 2.2988 min<sup>-1</sup> in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system, respectively. The results indicate that the rate of HO<sup>o</sup> generation is greatly enhanced by increasing Fe<sup>III</sup>-NTA concentration, especially in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system. This can be explained by the fact that high concentration of

Fe<sup>III</sup>-NTA would increase the generation rate of Fe<sup>II</sup>-NTA (R5) and (R7), which in turn reacts with  $H_2O_2$  to produce more HO.

#### 3.6. Intermediates and degradation pathways

It was worth noting that a high initial concentration of metoprolol (292  $\mu$ M, ten times higher than that in typical experiments) was used in order to capture more intermediates. The typical fullscan total molecular ion chromatograms of samples at different reaction time (RT) can be found in the Supplementary data. Two major peaks with molecular ion  $([M+H]^+)$  at mass-to-charge (m/z) ratios of 268 and 192 were observed in samples before the treatment, corresponding to metoprolol and NTA, respectively. Several intermediates were detected in samples after the reaction by mass spectrometry, their molecular ion were identified at m/z 134, 238, 258, 274, 282, 284, 298, 300, 314, 316, 330, 332 and 348. A major degradation product is 3-(isopropylamino)propane-1,2-diol (m/z134), which formed by the cleavage of aromatic ether bond of metoprolol due to the attack of HO. A same intermediate has been detected in the degradation of metoprolol by photocatalysis.  $\gamma$ -irradiation. Fenton and photo-Fenton processes [8,10,33,35]. In addition, the attack of HO on the ether bond and/or the  $\beta$ -C in methoxyethyl side chain of metoprolol gave intermediates with molecular ion at m/z 238, corresponding to 4-(2-hydroxy-3-(isopro pylamino)propoxy)benzaldehyde and 1-(4-ethylphenoxy)-3-(iso propylamino)propan-2-ol, consistent with previous studies [8,10].

The hydroxylation of benzene ring (and/or side chain) of metoprolol led to the formation of hydroxyl-metoprolol isomers (m/z 284). The intermediates with molecular ion at m/z 282 could be 4-(2-methoxyethyl)phenyl 2-hydroxy-3-(isopropylamino)pro panoate and 1-(4-(2-hydroxy-3-(isopropylamino)propoxy)phe nyl)-2-methoxyethanone, which were formed by the oxidation of hydroxyl group to keto group in hydroxyl-metoprolol (i.e., 3-(isopro pylamino)-1-(4-(2-methoxyethyl)phenoxy)propane-1,2-diol and 1 -(4-(1-hydroxy-2-methoxyethyl)phenoxy)-3-(isopropylamino)pro pan-2-ol). Moreover, the addition of 1 HO<sup>,</sup>, 2 HO<sup>,</sup>, 3 HO<sup>,</sup> and 4 HO<sup>,</sup> on hydroxyl-metoprolol gave intermediates with molecular ion at m/z300, 316, 332 and 348, corresponding to di-, tri-, tetra-, and pentahydroxyl-metoprolol, respectively. The intermediates with molecular ion at m/z 298 and 314 could be formed by the addition of 1 HO. and 2 HO<sup>•</sup> on the detected intermediates (m/z 282). The intermediates with molecular ion at m/z 258 and 274 were probably formed by the addition of 2 HO<sup>.</sup> and 3 HO<sup>.</sup> on two undetected intermediates (m/z 226), i.e., 1-amino-3-(4-(2-methoxyethyl)phenoxy)propan-2ol and 4-(2-hydroxy-3-(isopropylamino)propoxy)phenol [33,36].

A comparison of Figs. S2 and S3 demonstrates that Fe<sup>III</sup>-NTA/ H<sub>2</sub>O<sub>2</sub>-Mn system led to more efficient degradation of metoprolol and its intermediates. In addition, Fig. 8 shows the changes in the total molecular ion chromatogram area of metoprolol and its intermediates with the reaction time. It can be seen that the total areas of major intermediates (i.e., *m*/*z* 238, 282, 284, 298, 300, etc.) increased rapidly to reach their maximum values within the first 3–5 min of the reaction time, and then decreased to undetectable levels at 20 min of the reaction time. In particular, the major intermediate  $(m/z \ 134)$  increased continuously until almost complete degradation of metoprolol took place (12-15 min of the reaction time), and then began to decrease slowly over time. The degradation pathways of metoprolol were proposed on the basis of these identified intermediates, which were presented in Scheme 1. Our results clearly suggested that the degradation of metoprolol and its intermediates were mainly caused by HO' attack.

#### 4. Conclusion

The results of this study indicate that Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> is able to degrade metoprolol at initial neutral pH, and the degradation rate

of metoprolol can be greatly improved with the presence of  $Mn^{II}$ . The *k* values of metoprolol in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system were typically 7–9-fold larger than those in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> system. It can be concluded that the presence of  $Mn^{II}$  does not directly catalyze Fenton-like reaction, but indirectly enhance Fe<sup>III</sup>-NTA-catalyzed Fenton-like reaction over a wide pH range of 4.0–8.0. The main intermediates during the degradation of metoprolol have been identified by mass spectrometry, which suggested that the degradation of metoprolol and its intermediates were mainly caused by HO<sup>•</sup> attack. Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system led to more efficient degradation of metoprolol and its intermediates.

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#### Appendix A. Supplementary data

Quenching tests with 2-propanol and chloroform on the degradation of metoprolol in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn system (Fig. S1). Total molecular ion chromatogram of metoprolol and its intermediates in Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub> and Fe<sup>III</sup>-NTA/H<sub>2</sub>O<sub>2</sub>-Mn systems (Fig. S2 and Fig. S3). Supplementary data associated with this article can be found, in the online version, at http://dx.doi.org/10.1016/j.cej. 2016.12.098.

#### References

- M.S. Kostich, A.L. Batt, J.M. Lazorchak, Concentrations of prioritized pharmaceuticals in effluents from 50 large wastewater treatment plants in the US and implications for risk estimation, Environ. Pollut. 184 (2014) 354– 359.
- [2] J. Sun, Q. Luo, D. Wang, Z. Wang, Occurrences of pharmaceuticals in drinking water sources of major river watersheds, China, Ecotoxicol. Environ. Saf. 117 (2015) 132–140.
- [3] Q.-T. Liu, H.E. Williams, Kinetics and degradation products for direct photolysis of β-blockers in water, Environ. Sci. Technol. 41 (2007) 803–810.
- [4] C. Miège, J.M. Choubert, L. Ribeiro, M. Eusèbe, M. Coquery, Fate of pharmaceuticals and personal care products in wastewater treatment plants – conception of a database and first results, Environ. Pollut. 157 (2009) 1721– 1726.
- [5] C.-C. Chen, R.-J. Wu, Y.-Y. Tzeng, C.-S. Lu, Chemical oxidative degradation of acridine orange dye in aqueous solution by Fenton's reagent, J. Chin. Chem. Soc. 56 (2009) 1147–1155.
- [6] H.-J. Fan, S.-T. Huang, W.-H. Chung, J.-L. Jan, W.-Y. Lin, C.-C. Chen, Degradation pathways of crystal violet by Fenton and Fenton-like systems: condition optimization and intermediate separation and identification, J. Hazard. Mater. 171 (2009) 1032–1044.
- [7] C.-C. Chen, W.-C. Chen, M.-R. Chiou, S.-W. Chen, Y.Y. Chen, H.-J. Fan, Degradation of crystal violet by an FeGAC/H<sub>2</sub>O<sub>2</sub> process, J. Hazard. Mater. 196 (2011) 420–425.
- [8] B. Abramović, S. Kler, D. Šojić, M. Laušević, T. Radović, D. Vione, Photocatalytic degradation of metoprolol tartrate in suspensions of two TiO<sub>2</sub>-based photocatalysts with different surface area. Identification of intermediates and proposal of degradation pathways, J. Hazard. Mater. 198 (2011) 123–132.
- [9] F.J. Rivas, O. Gimeno, T. Borralho, M. Carbajo, UV-C radiation based methods for aqueous metoprolol elimination, J. Hazard. Mater. 179 (2010) 357–362.
- [10] V. Romero, S. Acevedo, P. Marco, J. Giménez, S. Esplugas, Enhancement of Fenton and photo-Fenton processes at initial circumneutral pH for the degradation of the β-blocker metoprolol, Water Res. 88 (2016) 449–457.
- [11] Y. Sun, J.J. Pignatello, Chemical treatment of pesticide wastes. Evaluation of iron(III) chelates for catalytic hydrogen peroxide oxidation of 2,4-D at circumneutral pH, J. Agric. Food Chem. 40 (1992) 322–327.
- [12] A. Rastogi, S.R. Al-Abed, D.D. Dionysiou, Effect of inorganic, synthetic and naturally occurring chelating agents on Fe(II) mediated advanced oxidation of chlorophenols, Water Res. 43 (2009) 684–694.
- [13] W. Huang, M. Brigante, F. Wu, C. Mousty, K. Hanna, G. Mailhot, Assessment of the Fe(III)–EDDS complex in Fenton-like processes: from the radical formation to the degradation of bisphenol A, Environ. Sci. Technol. 47 (2013) 1952–1959.

- [14] Y. Wu, M. Passananti, M. Brigante, W. Dong, G. Mailhot, Fe(III)–EDDS complex in Fenton and photo-Fenton processes: from the radical formation to the degradation of a target compound, Environ. Sci. Pollut. Res. 21 (2014) 12154– 12162.
- [15] Y. Zhang, N. Klamerth, S.A. Messele, P. Chelme-Ayala, M. Gamal El-Din, Kinetics study on the degradation of a model naphthenic acid by ethylenediamine-N, N'-disuccinic acid-modified Fenton process, J. Hazard. Mater. 318 (2016) 371– 378.
- [16] N. Klamerth, S. Malato, A. Agüera, A. Fernández-Alba, G. Mailhot, Treatment of municipal wastewater treatment plant effluents with modified photo-Fenton as a tertiary treatment for the degradation of micro pollutants and disinfection, Environ. Sci. Technol. 46 (2012) 2885–2892.
- [17] N. Klamerth, S. Malato, A. Agüera, A. Fernández-Alba, Photo-Fenton and modified photo-Fenton at neutral pH for the treatment of emerging contaminants in wastewater treatment plant effluents: a comparison, Water Res. 47 (2013) 833–840.
- [18] S. Miralles-Cuevas, I. Oller, J.A.S. Pérez, S. Malato, Removal of pharmaceuticals from MWTP effluent by nanofiltration and solar photo-Fenton using two different iron complexes at neutral pH, Water Res. 64 (2014) 23–31.
- [19] Y.H. Dao, J. De Laat, Hydroxyl radical involvement in the decomposition of hydrogen peroxide by ferrous and ferric-nitrilotriacetate complexes at neutral pH, Water Res. 45 (2011) 3309–3317.
- [20] S.-P. Sun, X. Zeng, A.T. Lemley, Kinetics and mechanism of carbamazepine degradation by a modified Fenton-like reaction with ferric-nitrilotriacetate complexes, J. Hazard. Mater. 252–253 (2013) 155–165.
- [21] Y. Zhang, N. Klamerth, M. Gamal El-Din, Degradation of a model naphthenic acid by nitrilotriacetic acid – modified Fenton process, Chem. Eng. J. 292 (2016) 340–347.
- [22] A. De Luca, R.F. Dantas, S. Esplugas, Assessment of iron chelates efficiency for photo-Fenton at neutral pH, Water Res. 61 (2014) 232-242.
- [23] Y. Li, J. Sun, S.-P. Sun, Mn<sup>2\*</sup>-mediated homogeneous Fenton-like reaction of Fe (III)-NTA complex for efficient degradation of organic contaminants under neutral conditions, J. Hazard. Mater. 313 (2016) 193–200.
- [24] C. Kormann, D.W. Bahnemann, M.R. Hoffmann, Photocatalytic production of hydrogen peroxides and organic peroxides in aqueous suspensions of titanium dioxide, zinc oxide, and desert sand, Environ. Sci. Technol. 22 (1988) 798–806.

- [25] R.J. Watts, J. Sarasa, F.J. Loge, A.L. Teel, Oxidative and reductive pathways in manganese-catalyzed Fenton's reactions, J. Environ. Eng. ASCE 131 (2005) 158–164.
- $\label{eq:26} \mbox{[26] B.H.J. Bielski, D.E. Cabelli, R.L. Arudi, A.B. Ross, Reactivity of HO_2/O_2^- radicals in aqueous solution, J. Phys. Chem. Ref. Data 14 (1985) 1041–1100.$
- [27] J. De Laat, Y.H. Dao, N. Hamdi El Najjar, C. Daou, Effect of some parameters on the rate of the catalysed decomposition of hydrogen peroxide by iron(III)nitrilotriacetate in water, Water Res. 45 (2011) 5654–5664.
- [28] G.V. Buxton, C.L. Greenstock, W.P. Helman, A.B. Ross, Critical-review of rate constants for reactions of hydrated electrons, hydrogen-atoms and hydroxyl radicals ('OH/'O<sup>-</sup>) in aqueous-solution, J. Phys. Chem. Ref. Data 17 (1988) 513– 886.
- [29] R.J. Motekaitis, A.E. Martell, The iron(III) and iron(II) complexes of nitrilotriacetic acid, J. Coord. Chem. 31 (1994) 67–78.
- [30] B.A. Smith, A.L. Teel, R.J. Watts, Identification of the reactive oxygen species responsible for carbon tetrachloride degradation in modified Fenton's systems, Environ. Sci. Technol. 38 (2004) 5465–5469.
- [31] M. Bucheli-Witschel, T. Egli, Environmental fate and microbial degradation of aminopolycarboxylic acids, FEMS Microbiol. Rev. 25 (2001) 69–106.
- [32] C. Slegers, G. Baldacchino, D. Le Parc, B. Hickel, B. Tilquin, Radical mechanisms in the radiosterilization of metoprolol tartrate solutions, Pharm. Res. 20 (2003) 1977–1983.
- [33] W. Song, W.J. Cooper, S.P. Mezyk, J. Greaves, B.M. Peake, Free radical destruction of β-blockers in aqueous solution, Environ. Sci. Technol. 42 (2008) 1256–1261.
- [34] C. Walling, Fenton's reagent revisited, Acc. Chem. Res. 8 (1975) 125-131.
- [35] H. Yang, T. An, G. Li, W. Song, W.J. Cooper, H. Luo, X. Guo, Photocatalytic degradation kinetics and mechanism of environmental pharmaceuticals in aqueous suspension of TiO<sub>2</sub>: a case of β-blockers, J. Hazard. Mater. 179 (2010) 834–839.
- [36] J. Radjenovic, B.I. Escher, K. Rabaey, Electrochemical degradation of the  $\beta$ -blocker metoprolol by Ti/Ru\_{0.7}lr\_{0.3}O\_2 and Ti/SnO\_2-Sb electrodes, Water Res. 45 (2011) 3205–3214.